

Effect of Endosulfan 35 EC on ATPases in the Tissues of a Freshwater Field Crab Barytelphusa guerini

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Pesticides are synthetic organic chemicals introduced intentionally in order to augment the agricultural productivity. But the injudicious and inadvertent use chemicals during the last three decades resulted in unprecedented chemical pollution, posing a threat to many nontarget organisms in aquatic Organochlorine insecticides are quite ecosystems. and consequently undergo biomagnification persistent and accumulate in the tissues (Wilber 1971: Gray and Peterle 1979). Endosulfan (1.4. 5, 6.7.7-hexachloro 8. 9. 10- Try-norbonn 5.2.3-enylene dimethyl sulphite) is broad - spectrum organochlorine insecticide, widely and is known for its potential for insect knockdown capacity. The deleterious effects of endosulfan on fish models have been well documented (Davis and Wedmeyer 1971: Ÿар et 1981). al. information on freshwater crustaceans, particularly on freshwater field crabs. that inhabit paddy fields, is very scanty. Barytelphusa guerini, a freshwater field of high edible importance to rural populations. received little attention with regard to endosulfan toxicity.

ATPases are the enzymes concerned with the immediate release of energy are useful for all types physiological activities. ATPase activity can be taken meaningful index of cellular activity and forms a toxicological tool. Hence the present investigation is aimed to understand the in vivo emulsifiable concentrate effect of an (35 EC) on ATPase system in the tissues of the field crab. Barytelphusa guerini.

MATERIALS AND METHODS

Healthy, uniform-sized (25.00±0.5) male crabs Barytelphusa guerini collected from and around *Correspondence to Dr.S.L.N.Reddy, Dept. of Zoology, Osmania University, Hyderabad-500 007, India.

Hyderabad were acclimated to laboratory conditions for a period of 15d. The animals were fed fish meat ad libitum. A density of 10 crabs per 8L of water was used with 10 individuals in each test container. The physico-chemical characteristics of the test water were as follows: pH 7.2-7.4; dissolved oxygen 7.8-8.0 mg/L; salinity 0.190 g/L; chlorinity 0.110 g/L; alkalinity 102 mg/L hardness of water 112 mg/L and carbon dioxide 2.08 mg/L. Endosulfan (Hoechest is a brownish crystalline solid, and stock India) solution was prepared in acetone and mixed in water in required dilutions. To determine the LCso value, the were exposed to six serial concentrations of endosulfan. The bioassay experiment of each concentration was repeated six times with parallel controls, and mortality was noted in each concentration at the end of 96 hr. The LCso value (17.78 mg/L for 96hr was determined according to Finney (1964).

The crabs were divided into two groups of 24 each. Group I served as control and group II were subsequently exposed to a sublethal concentration of endosulfan (1/3 of LC₅₀ i.e. 6 mg/L for a period of 4d. The toxicant water and control water were renewed every 24 hours after feeding. The crabs were starved day prior to experimentation to avoid metabolic differences, if any, due to differential feeding and Six crabs each from experimental and food reserves. control groups were sacrificed every 24 hr over a period of 4 days. The tissues, chelate leg muscle. hepatopancreas, heart, gills and thoracic ganglion were isolated from both control and toxicant-treated animals and were immediately transferred to a deepfreezer for enzyme assays, i.e., ATPases.

ATPase (Adenosine triphosphatase: ATPase phosphorylase E.C. 3.6.1.3) activity was assayed by the method of Kaplay (1978). The tissues were homogenized in 0.25M ice-cold sucrose solution containing 1mM EDTA and 30mM histidine (pH 7.4). The reaction mixture of the enzyme assay in a volume of 1 mL contained: 3mM ATP. 3mM MgCl2, 140 mM NaCl, 14mM KCl, 20mM Tris-HCl buffer (pH 7.4) and 0.2ml of ouabain (for Mg** ATPases). The reaction was initiated by adding ATP and allowed for 10 min at 37°C. The reaction was stopped by adding 1mL of 10% TCA. The inorganic phosphate liberated was by the method of Taussaky and Shorr (1953). measured The difference in the activity of enzyme in the absence and presence of ouabain was taken as Na* -K* ATPase activity. Total activity in the presence of ouabain was Mg**- ATPase activity. The protein content in the enzyme source was estimated according to Lowry et al. (1951), using bovine serum albumin (Sigma) as a standard. Enzyme activity was expressed as μ moles of inorganic phosphate formed/mg protein/hr. Students 't' test was used to compare the differences between control and experimental groups.

RESULTS AND DISCUSSION

ATPase activities measured in the present investigation to assess the toxicological effect of a sublethal concentration of endosulfan on the energy metabolism tissues of crab showed significant alterations in enzyme activities, reflecting sensitivity of crabs to endosulfan toxicity. ATPases, both Na* -K* ATPase and Mg** - ATPase, exhibited a progressive inhibition. Na+ - K+ ATPase is a key enzyme in cellular water balance and in osmoregulation in whole animal. It plays an important role in aquatic organisms. was almost an inhibition in the activity of Na* -K* The inhibition in the activity was, however. ATPase. tissue-specific (Table 1). However, a negligible enhancement in the enzyme activity in gills muscle on 1st day of exposure suggests the stimulation of anaerobic metabolism at the expense of aerobic There was a subsequent decrease in the Nat processes. -K* ATPase activity in all tissues during rest of the exposure. Several classes of pesticides were known inhibit ATPases activities (Koch 1969; 1970: Desaiah et al. 1972: Cutkomp et al 1976). ATPases are highly susceptible to organochlorine pesticides (Davis and Wedmeyer 1971). opinions were expressed for such an inhibition in the ATPase activity in the cellular system. Inhibition indicates an overall disruption in the Na' -K' ATPase enzyme is associated with metabolism. lipoprotein in the form of a complex (Nakao et al. 1974). Kinter et al. (1972) speculated lipophilic pesticidas exert biologia effect on ATPase system which would induce partitioning in the enzyme complex. The lipophilic partitioning of ATPase enzyme by endosulfan may produce allosteric change resulting decreased ATPase activity. Similar conclusion was drawn by La Rocca and Carlson (1979) for the inhibition of ATPase by PCB. Endosulfan toxicosis in fish models exhibited symptoms of hyperexcitabiliity, tremors, and convulsions (Singh and Srivastava 1981). These symptoms can be caused by the inhibition of Na* -K+ ATPases.

Oligomycin sensitive (mitochondrial) Mg** -ATPase in the present study exhibited a progressive inhibition when exposed to endosulfan. Similar observations were made in a teleost fish Channa gachua when exposed to endosulfan (Dalela et al. 1979). Mg** ATPases are associated with synthesis of ATP in

Table 1. Naº -K' ATPase activity in the tissues of Barytelphusa guerini during endosulfan toxicity

4 7 7 1 1					*************
GIIIS Control	rol	1.566 ± 0.032	1.588 ± 0.140	1.564 ± 0.026	1.592 ± 0.092
Expel	Experimental X Difference	1.046 I 0.086 + 5.23	1.550 I 0.054 - 2.01	1.554 I U.IIZ*** -14.70	1.245 I 0.096*** -21.73
Muscle Control	rol	2.274 ± 0.052	2.264 ± 0.068	2.288 ± 0.024	2.256 ± 0.126
Expe	Experimental	2.368 ± 0.120	2.122 ± 0.082*	1.992 ± 0.018***	1.818 ± 0.078**
* Dii	* Difference	+ 4.13	- 6.27	-12.93	-19.41
Hapato- Control	rol	1.658 ± 0.076	1.642 ± 0.048	1.644 ± 0.086	1.658 ± 0.084
pancreas Exper	Experimenta!	1.634 ± 0.022	1.572 ± 0.032	1.452 ± 0.064**	1.392 ± 0.068***
* Dil	* Difference	- 1.44	- 4.27	-11.676	-16.04
Heart Control	tol	2.394 ± 0.116	2.326 ± 0.112	2.432 ± 0.048	2.424 ± 0.142
Expe	Experimental	2.280 ± 0.042	2.144 ± 0.078**	2.122 ± 0.062***	2.152 ± 0.038***
iio *	* Difference	- 4.76	- 7.82	-12.74	-11.22
Thoracic- Control	rol	2.214 ± 0.038	2.242 ± 0.022	2.214 ± 0.084	2.258 ± 0.042
ganglion Exper	Experimental % Difference	2.138 ± 0.046 - 3.43	2.144 ± 0.014** - 4.37	1.964 ± 0.106***	2.008 ± 0.086***

0.998 ± 0.058*** .308 ± 0.084** 2.512 ± 0.054** 2.352 ± 0.064*** 1.204 ± 0.078** /alues expressed as µmoles of Pi/mg protein/hr. Each value is a mean of six observations ± SD. Experimental values 1.634 ± 0.028 2.968 ± 0.112 1.148 ± 0.146 2.664 ± 0.142 $.372 \pm 0.034$ -19.95 -11.04 Table 2. Mg.2 ATPase activity in the tissues of Barytelphusa guerini during endosulfan toxicity toxicity. 2.656 ± 0.054** 1.012 ± 0.074*** 1.462 t 0.108** 1.228 ± 0.054** 2.488 ± 0.104* 2.942 ± 0.102 1.134 ± 0.122 2.656 ± 0.068 1.344 ± 0.046 1.648 ± 0.062 - 6.32 -10.75 -11.28 - 9.72 2.788 ± 0.092*** 2.522 ± 0.052* 2,956 ± 0,036 ..524 ± 0.064 1.056 ± 0.104 2.634 ± 0.084 1.252 ± 0.062 1.622 ± 0.082 $.154 \pm 0.088$.358 ± 0.086 - 6.04 - 5.68 - 8.49 - 7.80 1.288 ± 0.084 - 2.862 2.602 ± 0.116 1.326 ± 0.158 2.908 ± 0.078 1.078 ± 0.098 2.558 ± 0.098 1.602 ± 0.058 1.548 ± 0.072 2.822 ± 0.056 1.112 ± 0.064 - 1.69 - 3.37 - 2.95 24hr ***************** % Difference * Difference xperimental bifference Experimental b) Ifference Experimenta! Experimental Experimental * Difference Control Control Control [{ssues Control Control **Fhoracic**ganglion pancreas Hapato-Musc le Heart Gills

significantly different from control with statistical significance at * P<0.001; ** P< 0.005; ***P < 0.001.

mitochondria oxidative phosphorylation. through Evidences also indicate implicating OS (mitochondrial) Mg** ATPases in the terminal step of oxidative phosphorylation result in the formation of Mitochondrial disorganization resulted during pesticidal toxicosis (Pardini et al. 1980) may cause inhibition of Mg** -ATPase in this study. It may be concluded that endosulfan induced inhibition in ATPases not only disrupt various aspects of cell function and also energy production necessary for maintenance of physiological processes. Further analysis of adenine nucleotides which are reliable estimates of adenylate energy charge (AEC) will provide needed insight into metabolic energy state of cells.

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REFERENCES

- Cutkomp LK, Desaiah D. Cheng EY. Ves EV. Koch RB (1976)
 The <u>in vivo</u> sensitivity of ATPase enzyme in cockroaches. Pestic Biochem Physiol 4: 232-238.
- Dalela RC. Bhatnagar MC. Tyagi AK. Verma SR (1979) Adenosine triphosphatase activity in few tissues of freshwater teleost <u>Channa gachua</u> following <u>in vivo</u> exposure to endosulfan. Toxicology 11:361-368
- Davis PW. Wedmeyer GA (1971) Na* -K* activated ATPase inhibition in rainbow trout. A site for organochlorine pesticide toxicity. Comp Biochem Physiol 40B: 823-827.
- Desaiah D. Cutkomp LK, Yap HH. Koch RB (1972) Inhibition of oligomycin-sensitive and insensitive Mg**-ATPase activity in fish by polychlorinated biphenyls. Biochem Pharmacol 21: 857-865
- Finney DJ (1964) Probit analysis A statistical treatment of the sigmoid response curve. Cambridge University Press. London
- Grav BL. Peterle W (1979) Compartmentalisation of C¹⁴ DDT in an experimental old field plot. Bull Environ Contam Toxicol 23: 846-853.
- Kaplay SS (1978) Erythrocyte membrane Na* -K* activated ATPase in protein-caloric malnutrition. Amer J Clin Nutr 31: 579-584
- Kinter WB. Merkens LS. Janiki RH. Guarino AM (1972) Studies on the mechanism of toxicity of DDT and PCBs: Disruption of osmoregulation in marine fish. Environ Hlth Pers 8: 169-173
- Koch RB (1969) Chlorinated hydrocarbon insecticides: Inhibition of rabbit brain ATPase activities. J. Neurochem 16: 269-271

- Koch RB (1970) Inhibition of animal tissue ATPase activities by chlorinated hydrocarbon pesticides. Chem Biol Interact 1: 199-209
- La Rocca PT. Carlson GP (1979) The effect of polychlorinated biphenyls on adenosine triphosphatase activity Toxicol Appl Pharmacol 48: 185-192
- Lowry OH, Rosebrough NJ, Farr AL, Randall RJ (1951) Protein estimation with Folin-phenol reagant. J Biol Chem 193: 265-275
- Nakao M. Nako T. Hara Y. Nagi F. Ygasaki S. Koi M. Nakagawa A. Kawai K (1974) Purification and properties of Na* -K* ATPase from guinea pig brain. Ann N Y Acad Sci 242: 24-33
- Pardini RS. Heidken JC. Baker TA. Payme B (1980) Toxicology of various pesticides and their decomposition products on mitochondrial electron transport. Arch Environ Contam Toxicol 9: 87-97
- Rao DMR. Devi AP, Murty AS (1981) Toxicity and metabolism of endosulfan and its effect on oxygen consumption and total nitrogen excretion of the fish Macrognathus aculeatum. Pestic Biochem Physiol 15: 282-287.
- Singh NN. Srivastava AK (1981) Effects of endosulfan on fish carbohydrate metabolism. Ecotoxicol Environ Safety. 5: 412-417.
- Taussaky HH. Shorr E (1953) A microcolorimetric method for the determination of inorganic phosphate. J Biol Chem 202: 675-681
- Wilber CG (1971) The biological aspects of water pollution 2nd Edn., Chales CT (Ed.), Springfield [1]
- Yap HH. Desaiah D. Cutkomp LK. Koch RB (1975) In vitro inhibition of fish brain ATPase activity by cyclodiene insecticide and related compounds. Bull Environ Contam Toxicol 14: 163-168.

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